



In-situ synthesis of heterojunction $\text{TiO}_2/\text{MnO}_2$ nanostructure with excellent performance in vacuum ultraviolet photocatalytic oxidation of toluene

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ABSTRACT

A series of $\text{TiO}_2/\text{MnO}_2$ heterojunction catalysts were fabricated through an in-situ hydrothermal method, and for the first time, catalysts with this heterojunction structure were used for VUV-PCO toluene degradation. After the MnO_2 coating, the $\text{TiO}_2/\text{MnO}_2$ catalysts performed excellently for both VOCs degradation and residual ozone decomposition with removal efficiency reaching 96.0% and 99.9%, respectively. The enhanced photocatalytic activity towards toluene degradation could be attributed to the heterojunction structure of $\text{TiO}_2/\text{MnO}_2$, which provides excellent contact between MnO_2 and TiO_2 , and suppresses the recombination of photogenerated electron-hole pairs. Moreover, benefiting from the MnO_2 , the ozone generated in the system could be efficiently utilized and eliminated during the VUV-PCO process. The photocatalytic mechanism of $\text{TiO}_2/\text{MnO}_2$ for the toluene degradation and ozone decomposition was proposed.

1. Introduction

Volatile organic compounds (VOCs) are known as one of the major hazardous air pollutants including different chemical compounds generated from anthropogenic and natural sources [1]. VOCs in the atmosphere may cause a series of threats to the environment, including the formation of photochemical smog, ozone, and atmospheric PM_{2.5}. In addition, long-term exposure to VOCs may also affect the health of human beings [2,3]. Indoor VOCs, such as formaldehyde, acetone, benzene and toluene, are associated with sick building syndromes including throat irritation, headache and fatigue. It has been proven that some of these airs also raise carcinogenic risk [4]. Therefore, it is crucial to develop efficient ways of removing the VOCs in our environment. Many technologies have been used to control the emission of VOCs, including thermal incineration, physical and chemical adsorption, biological treatment as well as photocatalytic oxidation [5–11]. Among those technologies, photocatalytic oxidation (PCO) of VOCs over metal oxide catalyst is a promising and a low-cost method of removing it at room temperature [12–17].

Recently, vacuum ultraviolet (VUV) photocatalytic oxidation was reported to perform excellently in the degradation of VOCs due to its unique oxidation process. Numerous reactive species of ozone (O_3) and hydroxyl radicals ($\cdot\text{OH}$) can be generated under the irradiation of

185 nm UV light. It has been demonstrated that these O_3 and $\cdot\text{OH}$ can efficiently decompose the organic intermediates during the reaction, which help reduce the accumulation of the by-products and avoid the deactivation of catalysts, resulting in an enhancement in VOCs degradation performance. Enormous catalysts have been explored for the PCO of VOCs, among which, TiO_2 is widely used due to its high activity, nontoxicity and low cost. It was reported that the photocatalytic activity of TiO_2 relies on both the surface atomic structure and exposure capability to reactive crystal facets

According to theoretical studies, the catalytic activity of {001} surface of anatase TiO_2 nanosheets (NSs) is much higher than that of the {101} surface [5]. The {001} surface of anatase TiO_2 NSs has a surface energy of about 0.98 J m^{-2} , while those of {101} and {100} are only about 0.49 J m^{-2} and 0.58 J m^{-2} , respectively. Yu et al. reported that anatase {001} and {101} facets exhibit different band structures and band edge positions [18]. Under the irradiation of sunlight, photogenerated holes will move from {001} to {101} facets, while electrons will move in the opposite pathway. Therefore, a surface heterojunction will form between the co-exposed {101} and {001} facets, which results in the improvement of the photocatalytic activity [6]. Wang et al. also reported that higher toluene adsorption capacity and degradation performance could be achieved when TiO_2 NSs have a higher ratio of {001} facets [19].

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However, as a by-product in the VUV process, excessive O_3 poses a health hazard that needs to be removed. Unfortunately, TiO_2 alone showed poor ability for O_3 removal, thereby, limiting its application in the VUV-PCO of VOCs. It was reported that O_3 could be decomposed efficiently in the presence of MnO_2 [13,14,20–25]. When modified with MnO_2 , $MnO_2/TiO_2/ZSM-5$ catalyst performed excellently during the utilization as well as elimination of O_3 . Nevertheless, the high recombination rate of photogenerated electrons (e^-) and holes (h^+), and the low decomposition efficiency of the by-products, resulting in a limited degradation efficiency of VOCs. Therefore, it is highly desirable to design and synthesize a multifunction catalyst with high efficiency of VOCs degradation as well as O_3 decomposition.

In the present work, MnO_2 was grown in-situ on the surface of TiO_2 [001] NSs through the reaction of $KMnO_4$ with a carbon layer intermediate. Toluene was chosen as the representative VOCs in this study due to its high toxicity. The catalysts were tested in a VUV-PCO system to evaluate their performance in toluene degradation and O_3 elimination. TiO_2 NSs with dominant high-energy {001} facets were fabricated first, a layer of carbon was then coated on the TiO_2 NSs to facilitate the in-situ growth of MnO_2 NSs as schematically depicted in Fig. S1. Combining with a thin layer of MnO_2 on the surface, this novel heterojunction structure of TiO_2/MnO_2 NSs performed excellently in the photocatalytic degradation of toluene, as well as exhibited a dual-function of O_3 utilization and destruction. For the first time, we report this heterojunction nanostructured TiO_2/MnO_2 NSs for the VUV photocatalytic degradation of toluene, which inspires a new approach to design multifunction catalysts that can improve the efficiency of toluene degradation.

2. Experimental

2.1. Materials

Titanium butoxide ($C_{16}H_{36}O_4Ti$) and D-(+)-Glucose monohydrate ($C_6H_{12}O_6 \cdot H_2O$) were purchased from Aladdin. Potassium permanganate ($KMnO_4$), Sodium sulfate (Na_2SO_4), hydrofluoric acid (HF) and ethanol (C_2H_5OH) were provided by Sigma Aldrich. All the chemicals used were without further purification and DI water ($18.2 \Omega \text{ cm}$) was used throughout this study.

2.2. Preparation of heterojunction photocatalysts

TiO_2 NSs with dominant {001} facets were fabricated by the hydrothermal method [18]. A total of 25 mL of TBOT was firstly put into a Teflon-lined autoclave, 4.5 mL of HF was then added to the TBOT solution dropwise under stirring for about 30 min. Thereafter the mixture solution was put into an oven and reacted at 180°C for 24 h. The obtained samples were washed with water and ethanol thrice, then dried at 60°C overnight, followed by calcination in air at 500°C for 2 h to remove the F in the TiO_2 NSs.

In preparing TiO_2/MnO_2 NSs, a layer of carbon was firstly coated on the surface of the TiO_2 NSs. 1 g TiO_2 NSs was added into 100 mL of glucose solution (0.04 M) and stirred overnight. After centrifugation and drying, the sample was heat-treated at 500°C in N_2 gas for 2 h. After the carbon deposition process, the sample was dissolved into 40 mL (0.005 M) $KMnO_4$ solution with drastic stirring for 30 min and kept at 120°C for 5 h in the oven for in-situ growth of MnO_2 . TiO_2/MnO_2 NSs were obtained after washing with distilled water and drying at 60°C for 12 h. The content of MnO_2 could be adjusted by changing the concentration of $KMnO_4$ solution during the synthesis process, a series of TiO_2/MnO_2 NSs with different percentage of Mn were synthesized and referred to as x-M-T ($x = 1, 2, 3, 4, 5$).

2.3. Material characterization

As-prepared samples were characterized by scanning electron microscopy (SEM; Hitachi S-4800, 15 kV) equipped with an energy dispersive X-ray spectrometer (EDX), transmission electron microscopy (TEM; Philips CM100), scanning transmission electron microscope (STEM, FEI Tecnai G2 20 S-TWIN), and powder X-ray diffraction (XRD) with Cu K α radiation at 40 kV and 40 mA (D8 Advance, Bruker). UV-vis diffuse reflectance spectra (UV-vis DRS) was determined on a Varian Cary 5000 Scan UV-vis-NIR spectrometer, with $BaSO_4$ as a reference. Raman spectroscopy was obtained using a HR Evolution Raman microscope (Labram HR800 Evolution) with a laser light of 633 nm. The valence band of Ti, Mn, and O of the as-prepared sample were determined by a Perkin Elmer PHI5000 X-ray photoelectron spectrometer with Al K α (1486.6 eV) radiation. Photoluminescence (PL) spectra were obtained by a Cary Eclipse Fluorescence Spectrophotometer. Fluorescence spectra were recorded by Edinburgh Instrument FLS920. The photoelectrochemical experiments were carried out using an electrochemical workstation (CHI660E, Shanghai Chenhua Ltd., China) in a

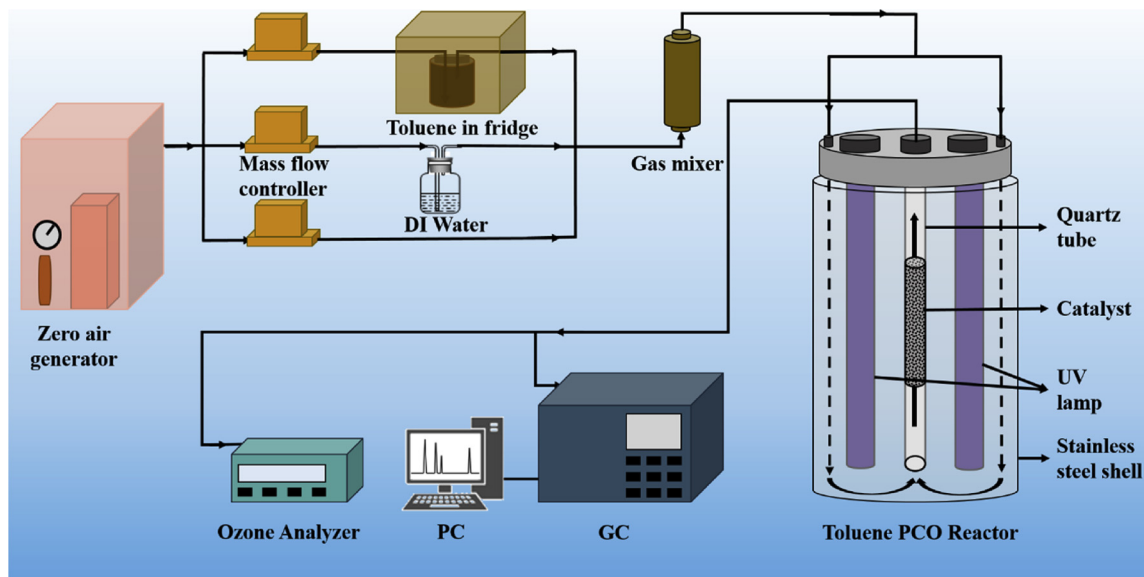


Fig. 1. Schematic diagram of the VUV-PCO set-up.

three-electrode cell, with 0.2 M Na₂SO₄ aqueous solution as electrolyte.

2.4. Photocatalytic activity test

Photocatalytic degradation of toluene was carried out in a continuous-flow fixed-bed reactor as shown in Fig. 1, which was reported in our previous study [24]. A quartz tube with a diameter of 9 mm was set in the middle of a stainless-steel reactor for catalysts loading, with two VUV lamps (4 W, Cnlight) on either side of the reactor. Toluene was injected into the reactor with an air stream at a flow rate of 1 L/min containing 30 ppm toluene, the humidity of the mixture gases was controlled at 50% (which has been proved to be the best in our previous study [26]). O₃ can reach its highest value (70 ppm) within 2 min under VUV lamps irradiation. The size of the catalysts was controlled at 20–40 mesh by a tablet press, with 1 g of the catalysts used for each test. The photocatalytic degradation tests were performed after purging for about 30 min to minimize the physical adsorption of toluene. The concentrations of outlet gases of toluene, CO and CO₂ at the outlet were detected every 10 min by a gas chromatograph (GC) (9790PLUS, Fuli Instruments, Zhejiang, China,) equipped with two flame-ionization detectors (FID with the detection range of toluene of 0.5 ~ 2500 ppm). The residual O₃ concentration was also examined by an ozone analyzer (202, 2 B Technology). All the experiments were carried out under the same conditions, and the total test time for each experiment was about 200 min. The catalyst was replaced by a fresh sample after each test cycle. Sample stability test was carried out using the same sample after recovering at 180 °C for 3 h in an oven. The toluene conversion and O₃ removal efficiency were calculated based on molar basis according to Eqs. (1) and (2), respectively.

$$\text{Toluene conversion \%} = \left(\frac{\text{Toluene}_{[\text{inlet}]} - \text{Toluene}_{[\text{outlet}]}}{\text{Toluene}_{[\text{inlet}]}} \right) * 100\% \quad (1)$$

$$\text{O}_3 \text{ removal efficiency \%} = \left(\frac{\text{O}_{3[\text{generate}]} - \text{O}_{3[\text{outlet}]}}{\text{O}_{3[\text{generate}]}} \right) * 100\% \quad (2)$$

3. Results and discussion

3.1. Structure and morphology

The phase structure and crystallinity of the as-synthesized samples were characterized by XRD and the corresponding phases are shown in Fig. S2. The characteristic peaks at 25.31°, 36.95°, 37.8°, 38.58°, 48.05°, 53.89°, 55.96°, 62.69°, 68.76°, 70.31°, 75.03° and 76.02° can be assigned to {101}, {103}, {004}, {112}, {200}, {105}, {211}, {204}, {116}, {220}, {215} and {301} facets of anatase TiO₂, respectively [27,28]. From the XRD patterns, we can see that all the samples are well-crystallized and the main diffraction of the peaks can be indexed to the anatase TiO₂ with tetragonal structure [JCPDS No. 73-1764] [18,29]. As shown in Fig. S2a, the broad {004} peak and the narrow {200} peak indicate that crystal growth is primarily limited to the [001] axis, resulting in the dominant {001} facet of TiO₂ nanosheets [30]. However, after coating with MnO₂, no significant MnO₂ peak can be observed in all the TiO₂/MnO₂ NSs samples (Fig. S2a). This may be due to the extremely low concentration and amorphous nature of the MnO₂, which was confirmed by EDS as shown in Table S1. However, the diffraction peaks ascribed to anatase TiO₂ phase are weaker with increase in MnO₂ content (Fig. S2b). According to the calculation, the crystallite size of TiO₂ NSs shows a little decline from 28.5 nm of TiO₂ NSs to 25.2 nm of 5-M-T, reflecting that the presence of MnO₂ prevents the crystallization of TiO₂ nanosheets [31].

To investigate the surface compositions and chemical states of the TiO₂/MnO₂ NSs (3-M-T), XPS was carried out as shown in Fig. 2, which exhibits main peaks of O 1s, Mn 2p and Ti 2p on the 3-M-T sample, indicating the presence of Mn. The main O 1s peak position at about

529.40 eV corresponds to lattice oxygen of the Ti-O bond [20]. The Mn 2p XPS spectra exhibits two peaks at about 641.67 eV and 653.78 eV, corresponding to the Mn2p_{3/2} and Mn2p_{1/2}, respectively [32]. The ratio of Mn³⁺/Mn⁴⁺ was about 0.6 calculated from the areas of the two peaks of Mn 2p, indicating that both MnO₂ and Mn₂O₃ exist in the samples. Since MnO₂ has a large proportion and some reports in literature regarded the material as MnO₂ [33–35], therefore, we interpreted the heterojunction catalyst as TiO₂/MnO₂ in our manuscript. It was reported that the existence of Mn³⁺ will lead to the generation of oxygen vacancies for maintaining electrostatic balance, which can act as active sites during the photocatalytic oxidation process [36–38]. The higher ratio of Mn³⁺/Mn⁴⁺ in the sample indicates the higher content of oxygen vacancies. Thus, this high ratio of Mn³⁺/Mn⁴⁺ (about 0.6) in the present sample indicates a large amount of oxygen vacancies of TiO₂/MnO₂ NSs, which also means a higher activity for the oxidation of pollutants. In Ti 2p region, the Ti 2p_{1/2} and Ti 2p_{3/2} main peak positions of TiO₂/MnO₂ NSs occur at 458.83 eV and 464.50 eV, respectively (Fig. 2d).

The morphologies of the as-synthesized samples were characterized by TEM as shown in Fig. 3. The well-defined TiO₂ NSs shows a rectangular outline structure with a side length of about 40 nm and a thickness of about 6 nm, which is in agreement with the report of Yu et al. [18]. After coating with MnO₂, some differences can be observed on the TiO₂ NSs. When the concentration of KMnO₄ is very low, almost no MnO₂ can be detected in the TEM image of the 1-M-T sample (Fig. 3b). However, when the concentration of KMnO₄ solution increased, a little amount of MnO₂ can be found in the 2-M-T sample (Fig. 3c). Further increase in the concentration of KMnO₄ solution leads to the coating of sheet structured MnO₂ on the surface of TiO₂ (001) NSs (Fig. 3d and e). These MnO₂ nanosheets are very thin and can be easily distinguished from TiO₂ (001) NSs. Further increase in the concentration of KMnO₄ leads to a corresponding coating of MnO₂ NSs on the surface of TiO₂. However, when MnO₂ reached a certain level, MnO₂ NSs merged together and grew into large particles, as can be seen in Fig. 3f. In the 4-M-T sample, almost all the TiO₂ (001) NSs are coated with MnO₂ NSs, and the excessive MnO₂ NSs appeared in large particles. The contents of MnO₂ NSs in the TiO₂/MnO₂ NSs samples were detected by the EDS as shown in Table S1. Only about 0.15 wt.% of Mn was observed in the 1-M-T sample, which increased with further increase in KMnO₄ concentration during the synthesis and reached about 1.59 wt.% for the 5-M-T sample.

The SEM images of TiO₂ and the 3-M-T sample are shown in Fig. S3 displaying the morphology of the catalysts. These two catalysts showed nanosheet structures with relatively large thickness, which may be due to the calcination for removing the F ion. However, after coating with MnO₂, some thin nanosheets can be observed on the surface of TiO₂ NSs, which can be assigned to MnO₂ NSs and agrees with the TEM images mentioned above. To acquire more information about the morphology of the photocatalyst, high-resolution TEM (HRTEM) and mapping were performed on the 3-M-T sample (Fig. 4). The side view of TiO₂ NSs is shown in Fig. 4b and the inset figure displays a d-spacing of about 0.235 nm, corresponding to (004) anatase crystalline phase that agrees well with the literature [19,39]. The TEM-EDX mapping of the 3-M-T sample shows the presence of Ti, O and Mn elements, indicating uniform distribution of Mn on the TiO₂ NSs.

3.2. Optical properties

UV-vis absorption spectra obtained from the TiO₂ NSs and TiO₂/MnO₂ NSs with different loading amounts are compared and shown in Fig. 5a. The as-prepared TiO₂ NSs displayed an absorption edge of about 410 nm. After coating with MnO₂, all the TiO₂/MnO₂ NSs samples showed a decline in the area of UV region absorption (about 200–340 nm). This may be due to the fact that the MnO₂ NSs prevents the light absorption of TiO₂. However, the absorption range of TiO₂/MnO₂ NSs was extended from 410 nm to near infrared light (800 nm).

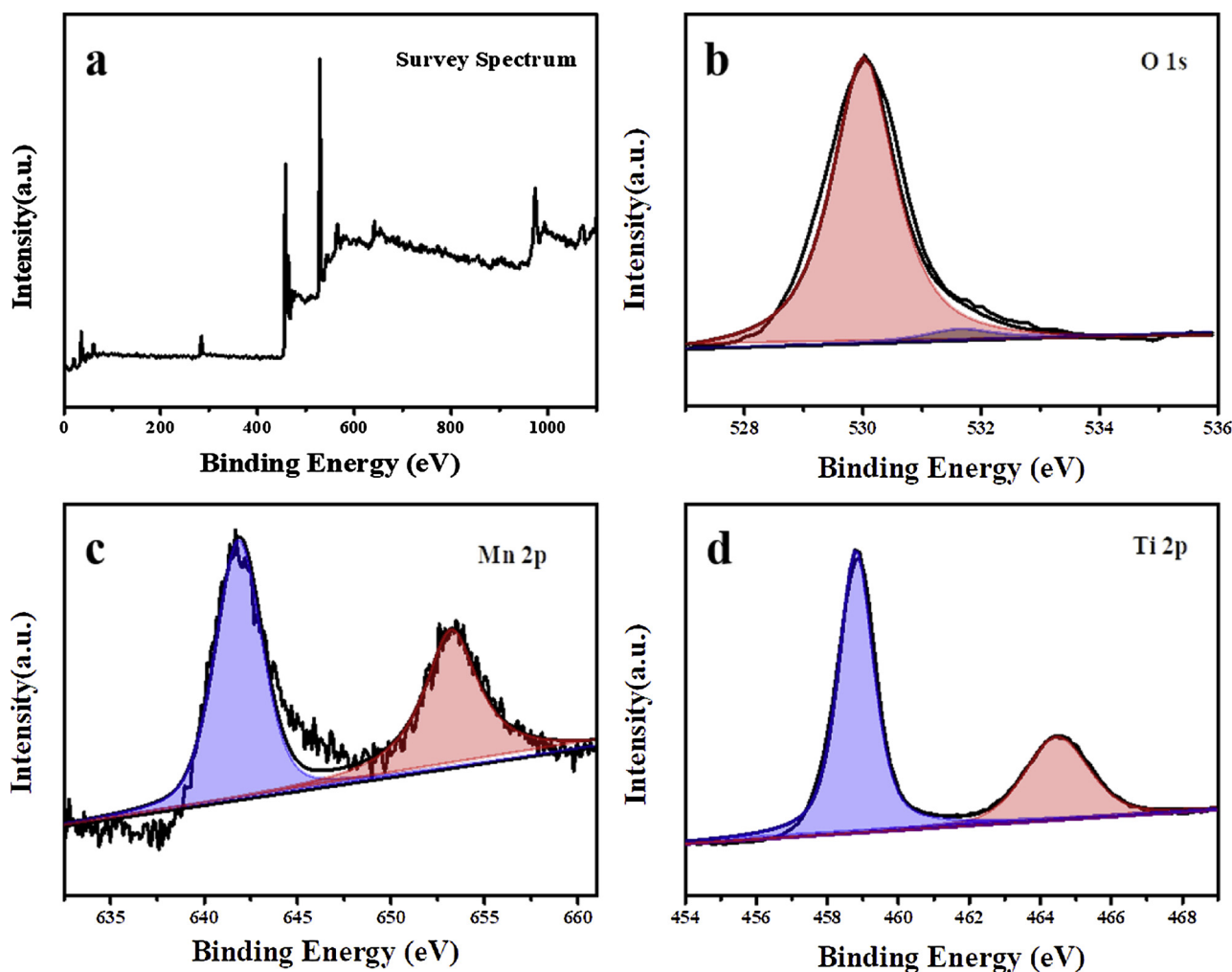


Fig. 2. (a) XPS spectrum of the as-prepared samples of 3-M-T, XPS spectra of (b) O 1s, (c) Mn 2p and (d) Ti 2p peaks of 3-M-T.

This significant extension of light absorption may be due to the narrow bandgap of MnO_2 [40]. In addition, light absorption intensity of $\text{TiO}_2/\text{MnO}_2$ NSs in visible light region increased with increase in the proportion of MnO_2 . The band gap energy (E_g) of the samples were calculated by the Kubelka-Munk method (Eq. 3) as follows: [41]

$$\alpha h\nu = A(h\nu - E_g)^2 \quad (3)$$

As shown in Fig. 5b, the E_g value of TiO_2 is about 3.0 eV. After coating with MnO_2 , the E_g value of the 3-M-T sample is about 2.25 eV, confirming the re-shift in the $\text{TiO}_2/\text{MnO}_2$ heterojunction compared with pristine TiO_2 NSs.

The phase compositions of the TiO_2 and $\text{TiO}_2/\text{MnO}_2$ NSs were investigated by Raman spectroscopy. Fig. 5c shows the fundamental Raman spectra of the samples. The 4 main peaks, observed at 145 cm^{-1} (E_g mode), 396 cm^{-1} (B_{1g} mode), 518 cm^{-1} (A_{1g} mode), and 640 cm^{-1} (E_g mode), were all due to TiO_2 [42]. The peak observed at 572 cm^{-1} was attributed to Mn-O stretching in the basal plane of MnO_6 sheets [43,44]. It can be seen that, after coating with MnO_2 nanosheets, the intensity of the E_g peaks at 145 cm^{-1} and 640 cm^{-1} decreased slightly, while the intensity of both the B_{1g} peak (at 396 cm^{-1}) and A_{1g} peak (at 518 cm^{-1}) showed a significant reduction. It is reported that the E_g peak is mainly caused by symmetric stretching vibration of O-Ti-O in TiO_2 , the B_{1g} peak is caused by symmetric bending vibration of O-Ti-O, while the A_{1g} peak is caused by antisymmetric bending vibration of O-Ti-O [23]. The percentage of the exposed anatase TiO_2 (001) facets has a direct relationship with the intensity variations of the Raman vibrational mode of the E_g and A_{1g} peaks [44].

PL spectra of the samples were obtained to investigate the carriers' separation efficiency. As shown in Fig. S4a, the PL quantum yield of the $\text{TiO}_2/\text{MnO}_2$ heterojunction was greatly reduced compared with that of the pristine TiO_2 , demonstrating an enhanced charge carriers' separation and reduced recombination. The time-based fluorescence spectra further confirmed the efficient charge carriers' separation over the 3-M-T sample (Fig. S4b). The average lifetime of the carriers (τ_{avg}) is about 2.07 ns for the 3-M-T sample, which is much longer than that of the pristine TiO_2 (about 0.69 ns). The increased τ_{avg} can be attributed to the formation of heterojunction that prolong the lifetime of the photo-generated charges.

3.3. Electrochemistry properties

To analyze the separation and migration efficiency of the photo-generated electrons and holes of the samples, photoelectrochemical measurements were carried out using an electrochemical workstation and the results are shown in Fig. 6. Typical i-t curves for the as-prepared electrodes of TiO_2 and 3-M-T samples were obtained using a 0.2 M Na_2SO_4 solution as electrolyte at a constant bias voltage of 0.8 V vs. SCE under the irradiation of two VUV lamps. The photocurrent density of pristine TiO_2 NSs is about $3.8\text{ }\mu\text{A}/\text{cm}^2$, while the photocurrent density of the mixture of TiO_2 NSs with MnO_2 is slightly higher than that of the pristine TiO_2 . However, after coating with MnO_2 , the photocurrent density of the 3-M-T sample increased to about $6.2\text{ }\mu\text{A}/\text{cm}^2$ under the same condition (Fig. 6a). This enhancement of photocurrent density indicated the efficient separation promotion of

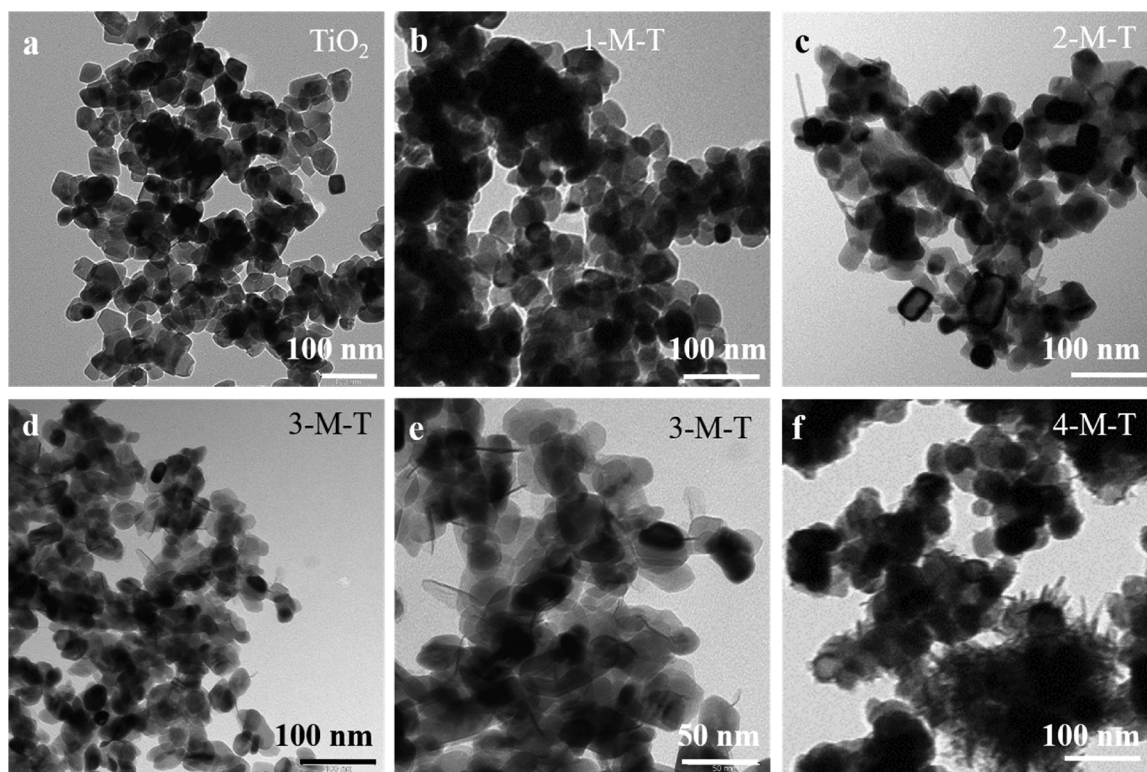


Fig. 3. TEM images of TiO₂ NSs and TiO₂/MnO₂ NSs.

photogenerated charge carriers and inhibition of the recombination of electron-hole pairs. After coating with MnO₂, heterojunction structure was constructed over the TiO₂/MnO₂ sample, which enhanced the transfer of photogenerated charge carriers. Under the irradiation of VUV light, the electrons generated by the excited state of TiO₂ will be transferred to the conduction band of MnO₂, resulting in an effective separation and migration of photogenerated electron-hole pairs [40].

The electrochemical impedance spectroscopy (EIS) measurements were carried out to study the interfacial properties of the photocatalysts as shown in Fig. 6b. The EIS was conducted in the frequency range of 10⁴–10 Hz using an amplitude of 5 mV at the open circuit potential of the system. The semicircle of the Nyquist plots at high frequencies showed the characteristic of the charge transfer process of the photocatalysts. Smaller diameters indicate lower charge-transfer resistance of the sample. It is clear that the 3-M-T heterojunction catalyst exhibited the smallest diameter among all the samples, indicating the lowest charge transfer resistance, and hence, the lowest recombination rate of the photogenerated electron-hole pairs [18,45].

Mott-Schottky measurements were carried out to investigate the electronic structure of the as-prepared samples. The Mott-Schottky plots of the samples were measured at 1000 Hz versus Ag/AgCl electrode. The carrier density (N_D) of the samples can be obtained by the following Mott-Schottky Eq.

$$\frac{1}{C^2} = \frac{2}{\epsilon\epsilon_r\epsilon_0 N_D A^2} \left(E - E_{fb} - \frac{kT}{e} \right) \quad (4)$$

where C is the space charge capacitance, E is the applied potential, e is the electron charge, k is the Boltzmann constant, T is the absolute temperature, A is the active area of the electrode, E_{fb} is the flat-band potential, ϵ_r and ϵ_0 are the semiconductor and free space dielectric constant, respectively [46,47]. As shown in Fig. 6c, the E_{fb} for the TiO₂ and 3-M-T are -0.51 V and -0.58 V (vs. Ag/AgCl), respectively. The N_D calculated from the slope of the linear region for TiO₂ and 3-M-T samples are $5.27 \times 10^{18} \text{ cm}^{-3}$ and $6.31 \times 10^{18} \text{ cm}^{-3}$, respectively. The higher carrier density of the 3-M-T sample than that of pristine TiO₂

sample indicates the enhanced charge transport. This result confirms that the formation of heterojunction in the 3-M-T sample could result in a higher photocatalytic activity in VOCs degradation.

3.4. Photocatalytic activity

Toluene removal efficiencies over different as-synthesized samples under the VUV-PCO process are shown in Fig. 7. As can be seen, the photolysis efficiency of toluene degradation in the VUV-PCO system is about 69.4%, which increased to 77.8% with the additional use of TiO₂ NSs photocatalysts due to the catalytic effect of TiO₂ under light irradiation. However, with increase in testing time, the removal efficiency of TiO₂ NSs showed a decline from 82.5% to about 76.0%, which may be due to the accumulation of by-products on the surface of TiO₂ [40]. However, after coating with MnO₂, the toluene removal efficiency of the 1-M-T heterojunction catalyst reached 93.5%. Moreover, when the content of Mn in the composite catalysts increased from 0.15% (1-M-T) to 0.67% (3-M-T), the toluene removal efficiency reached its highest value of about 96.0% (3-M-T). This excellent performance of the TiO₂/MnO₂ NSs illustrated that the introduction of MnO₂ during the VUV-PCO process can help to decompose the by-products on the surface of TiO₂. On the other hand, MnO₂ on the photocatalysts surface can also effectively utilize the O₃ produced by the VUV light irradiation to further decompose the toluene in the exhaust. However, when the content of Mn further increased, i.e. 1.19% (4-M-T) and 1.75% (5-M-T), the toluene removal efficiency dropped to about 89.3% and 85.2%, respectively. This means that the content of MnO₂ has the optimum value, at about 0.67% for the TiO₂/MnO₂ NSs catalysts. The lower activity at the high MnO₂ level may be due to the fact that too much MnO₂ coated on the surface of TiO₂ NSs will reduce the light absorption and block the active sites of TiO₂, leading to a lower photocatalytic activity for the degradation of toluene over TiO₂.

To investigate the degree of mineralization, the outlet CO_x concentration generated during the VUV-PCO process was measured by GC. The concentration of CO and CO₂ generated from toluene degradation

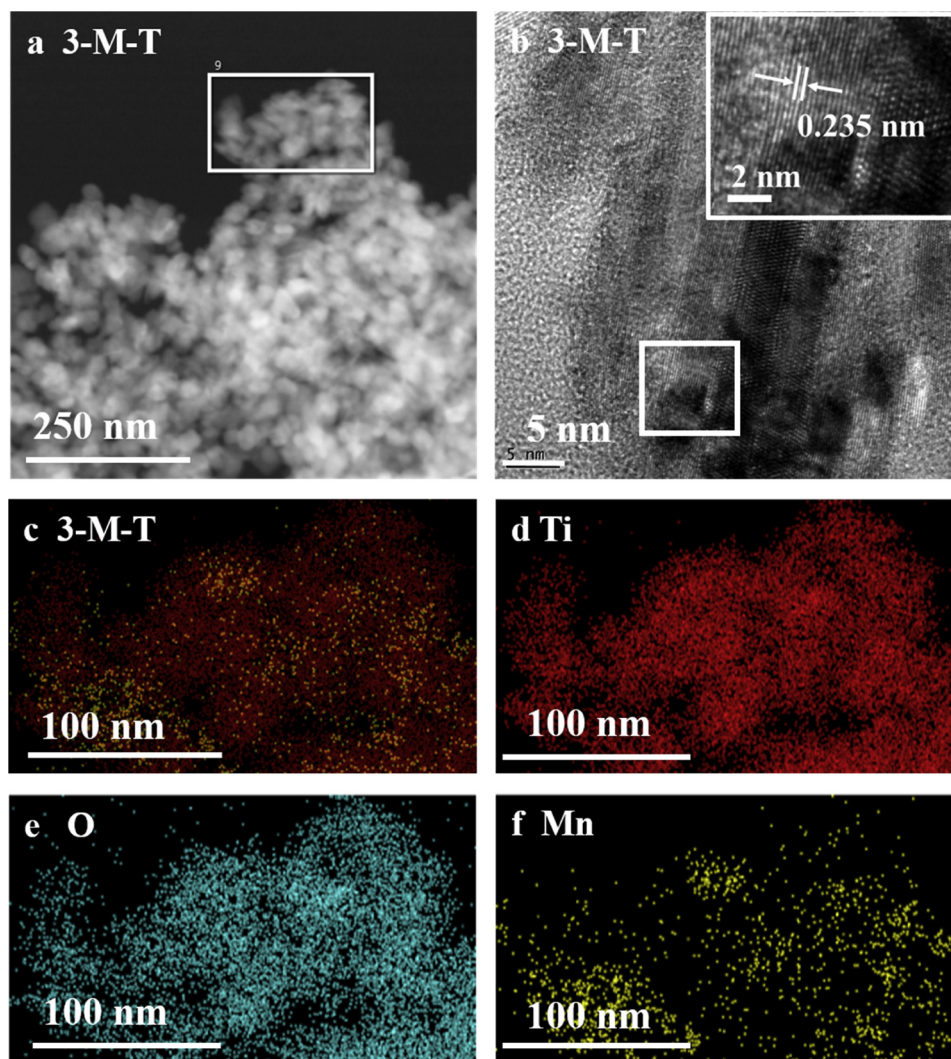


Fig. 4. (a,b) HRTEM images of 3-M-T and (c–f) corresponding mapping images of Ti, O and Mn.

over different catalysts were shown in Fig. 8. Both CO_2 and CO can be detected after the degradation of toluene. The concentration of CO and CO_2 for pristine TiO_2 was about 23 ppm and 45 ppm, respectively, which is a little lower than that of the simply photolysis process. This may be due to the by-products accumulation on the surface of TiO_2 because of its weak ability to decompose them. After coating with MnO_2 , the CO concentration showed an increase over all the $\text{TiO}_2/\text{MnO}_2$ NSs catalysts. It reached 34 ppm after coating a small amount of MnO_2 over the 1-M-T sample ($\text{Mn} = 0.15 \text{ wt.}\%$). However, the CO

concentration showed a decreasing trend with increase in the content of MnO_2 , and decreased to about 27 ppm for the 5-M-T sample ($\text{Mn} = 1.75 \text{ wt.}\%$). On the contrary, the outlet CO_2 generated from toluene degradation showed a significant improvement after coating with MnO_2 . At first, the CO_2 concentration increased with increasing content of MnO_2 in the $\text{TiO}_2/\text{MnO}_2$ NSs. However, after reaching its maximum at about 74 ppm (3-M-T and 4-M-T), the CO_2 generated from toluene degradation decreased to about 69 ppm for the 5-M-T sample, showing a similar trend as the toluene degradation (Fig. 8a). In conclusion,

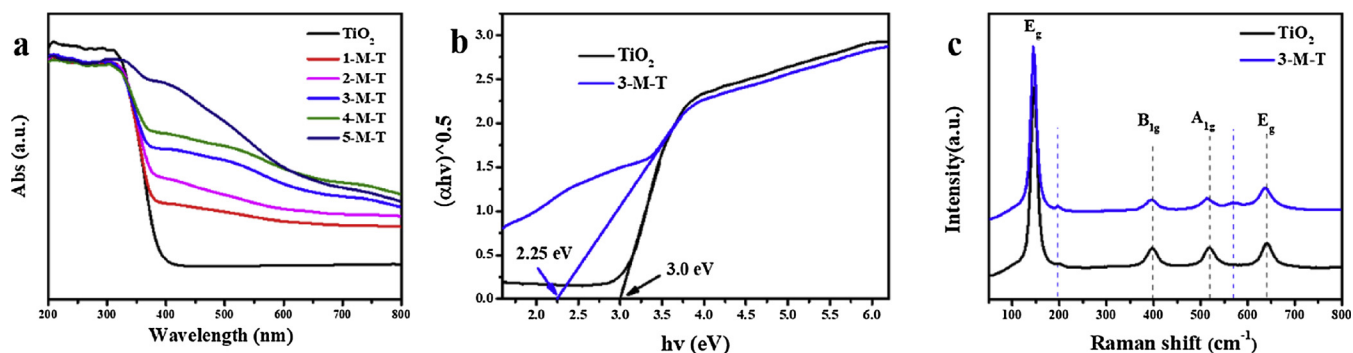


Fig. 5. (a) UV-vis absorbance spectra patterns of as-prepared samples, (b) the plot of $(\alpha h\nu)^{0.5}$ vs. energy $h\nu$ and energy band gap of TiO_2 and 3-M-T, (c) Raman shift of TiO_2 and 3-M-T.

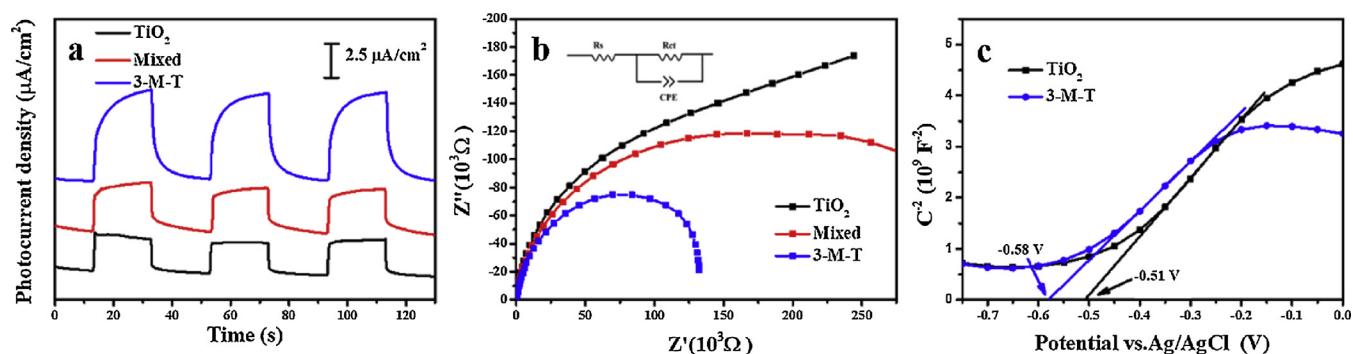


Fig. 6. Electrochemical properties of as-prepared samples, (a) Photocurrent density of TiO_2 , mixed (TiO_2 NSs and MnO_2) and 3-M-T, (b) Nyquist plots and (c) Mott-Schottky plots of TiO_2 and 3-M-T.

owing to the catalytic ozonation effect provided by MnO_2 and the synergistic effect of VUV-PCO, this heterojunction $\text{TiO}_2/\text{MnO}_2$ nanostructure composites showed an excellent performance for toluene degradation and mineralization.

To investigate the effect of O_3 on toluene degradation in the VUV-PCO process, we measured the residual O_3 concentration over different catalysts using an ozone analyzer and the result is shown in Fig. 9. Under the irradiation of VUV light, the O_3 generated in the system was regarded to be the same since the same VUV lamps were used for all the tests. First, we measured the O_3 concentration over the photolysis, which was used as a standard to calculate the O_3 removal efficiency for all the catalysts. Previous work has proved that irradiated TiO_2 can act as an active center for O_3 decomposition into O_2 and O^- [48]. As shown in Fig. 9, the O_3 removal efficiency of TiO_2 (001) NSs was only about 11.7%, showing its poor ability to decompose O_3 . However, after coating with MnO_2 , almost all the residual O_3 has been decomposed over the $\text{TiO}_2/\text{MnO}_2$ heterojunction catalysts. All the O_3 decomposition efficiency over $\text{TiO}_2/\text{MnO}_2$ NSs with different Mn percentage were about 99.7%, indicating that the introduction of MnO_2 can greatly promote the O_3 decomposition, which can be due to the active sites provided by Mn for catalytic oxidation of O_3 . In the VUV-PCO process, both the O_3 decomposition and toluene degradation have been improved after the coating of MnO_2 , which means that the Mn is crucial for toluene degradation via ozone-assisted catalytic oxidation. Therefore, $\text{TiO}_2/\text{MnO}_2$ heterojunction catalysts exhibited multi-functions in the VUV-PCO system. The toluene removal efficiency can reach 96% and almost all the O_3 by-product can be effectively utilized or eliminated during the toluene degradation process. Table 1 summarized the literature data on VOCs photocatalytic degradation over various TiO_2 -based catalysts [26,49–54].

In order to test the stability of $\text{TiO}_2/\text{MnO}_2$ NSs, we increased the testing time of the VOCs degradation. As shown in Fig. 10a, the VOCs removal efficiency was about 95% at the beginning, which was slightly raised to about 96% after 3 h testing, and remained rather stable after 6 h. Thereafter, we investigated the regeneration ability of this catalyst and the result is shown in Fig. 10b. Calcination was carried out at 180°C for 3 h to remove the by-products on the surface of the catalyst after the 3 h testing, and the same catalyst was used for the toluene degradation test. As shown in Fig. 10b, the 3-M-T sample showed a stable toluene removal efficiency during the whole 5-cycle tests. Even though it showed a slight fluctuation, the toluene removal efficiency was still maintained at about 96% throughout the 5 cycles tested. This result indicates that the $\text{TiO}_2/\text{MnO}_2$ NSs has a high stability for the VUV-PCO of toluene.

To confirm the advantages of this heterojunction structure of the $\text{TiO}_2/\text{MnO}_2$ heterojunction catalysts, comparison experiments have been conducted as shown in Fig. 11a. The performance of $\text{TiO}_2/\text{MnO}_2$ heterojunction, and a mixture of MnO_2 nanoparticles and the TiO_2 NSs with the same amount of MnO_2 were evaluated. As shown in the figure, the toluene removal efficiency of the mixture is only about 83%, which is much lower than that of the $\text{TiO}_2/\text{MnO}_2$ heterojunction (about 96%), indicating that the heterojunction structure of the catalyst promoted the catalytic efficiency in the VUV-PCO toluene degradation due to the increased contact area between TiO_2 and MnO_2 interface. As demonstrated in Fig. 6, the heterojunction structure facilitates the separation and migration of the photogenerated electrons and holes, thus, enhancing the photocatalytic activity of toluene degradation.

To evaluate the effect of hydroxyl radicals in the degradation of toluene under the VUV-PCO process, tert-butanol (TBA) was used as a scavenger for hydroxyl radicals ($\cdot\text{OH}$). 10 mM TBA was added dropwise

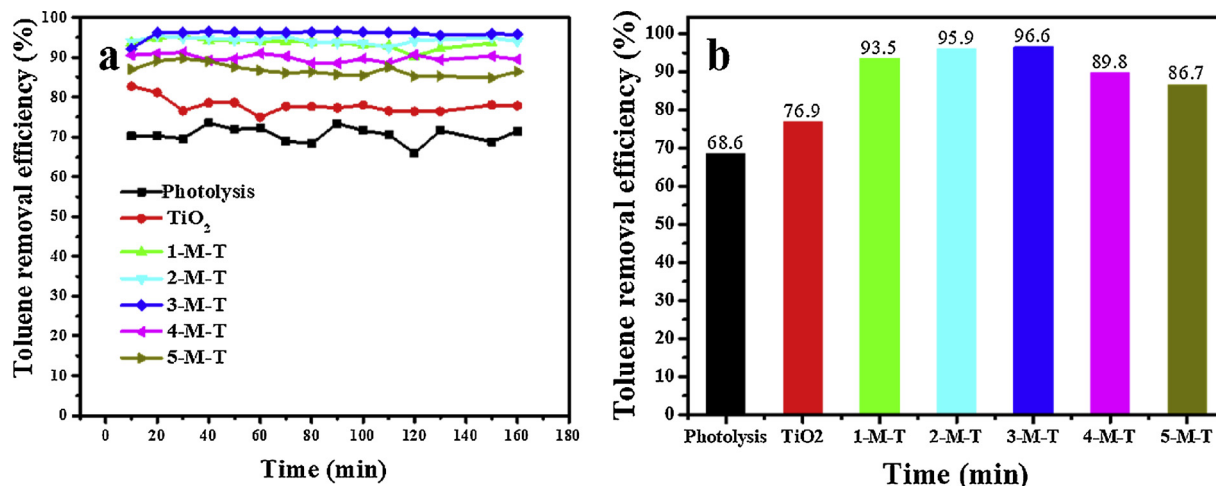


Fig. 7. Photocatalytic degradation of Toluene performance of Photolysis, TiO_2 NSs and $\text{TiO}_2/\text{MnO}_2$ NAs.

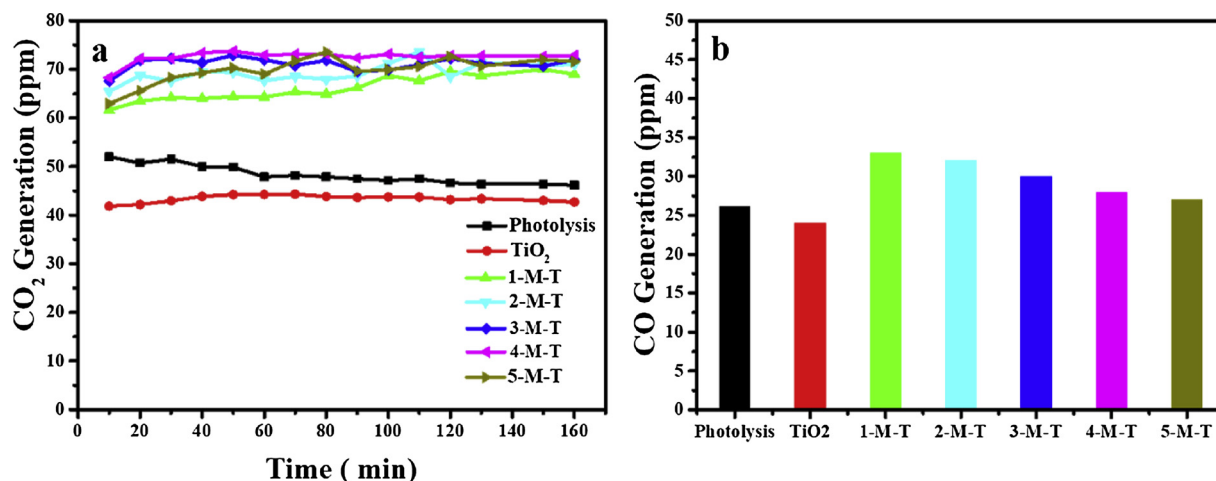


Fig. 8. Outlet CO₂ and CO concentration of different catalysts in VUV-PCO of Toluene.

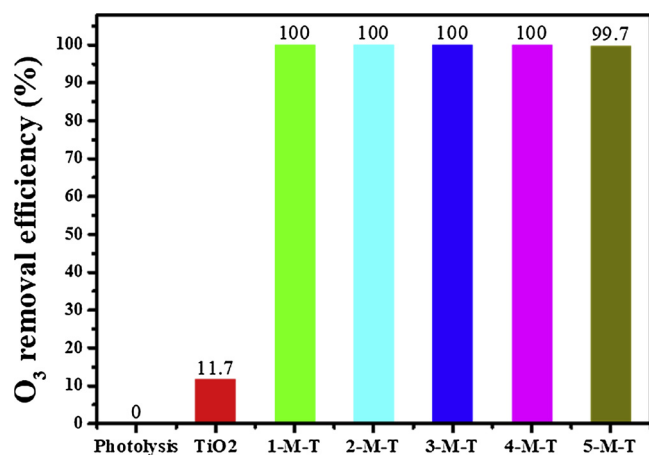


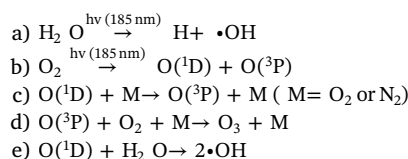
Fig. 9. O₃ removal efficiency of TiO₂ NSs and TiO₂/MnO₂ NSs in the VUV-PCO system.

to the as-prepared 3-M-T sample, and then toluene degradation performance of the catalyst was measured. The results of the toluene removal of the 3-M-T sample and the TBA treated 3-M-T sample are shown in Fig. 11b. It was found that the toluene removal efficiency of the 3-M-T sample decreased from about 96% to 80% after the treatment with TBA, which means the presence of TBA had a negative effect on the degradation of toluene. This result indicated that hydroxyl radicals has critical effect on the VUV-PCO toluene degradation system, which is in-line with our previous study [24].

3.5. Scheme of toluene degradation in VUV-PCO process

Compared with PCO, the VUV-PCO process provides more pathways for the toluene degradation. Under the irradiation of VUV lamp, all of

the PCO, photolysis and ozone catalytic oxidation (OZCO) can be well complemented with each other and have a synergistic effect for toluene oxidation. In the VUV-PCO system, besides the UVC (245 nm), there are also VUV (185 nm) irradiation from the VUV lamp that can generate hydroxyl radicals ([•]OH) and ozone. Under the irradiation of 185 nm VUV light, the oxygen and water in the air will react with the energetic photons and dissociate into oxygen species, including [•]OH, O(¹P) and O(³P). Then, the oxygen in the air can react with O(³P) species to generate ozone via the following steps (a–e) [20,21,55,56].



According to Fig. 6, the introduction of MnO₂ also accelerates the separation of electron-hole pairs, which enhances the toluene degradation. TiO₂ had limited effect on O₃ decomposition. However, after coating with MnO₂, TiO₂/MnO₂ heterojunction catalyst performed excellently in the elimination of O₃, indicating that O₃ generated in the first step could be utilized and decomposed on the active center of MnO₂ via ozone catalytic oxidation of toluene. Taking all the above results into consideration, the possible photocatalytic mechanism of toluene degradation in the VUV-PCO process over TiO₂/MnO₂ heterojunction is proposed as shown in Fig. 12. Under the irradiation of the VUV light, electrons and holes can be generated on the TiO₂ semiconductor. Based on the band energy structure, the conduction band (CB) position of MnO₂ is higher than that of TiO₂ (−0.29 eV). Therefore, under the irradiation of VUV light, the photo-induced electrons of TiO₂ can be transferred to the CB of MnO₂. However, O₂ cannot be reduced to [•]O₂[−] (−0.33 eV) due to the positive CB edge position of MnO₂ (1.33 eV) [57]. Fortunately, the O₃ generated in the first step can act as

Table 1
Literature data on VOCs photocatalytic degradation over various TiO₂-based catalysts.

Catalyst	Light source	Target pollutant	Concentration	Efficiency	Reference
Pd-TiO ₂	UV-LEDs 100 W	Benzylic alcohols	0.1mmol	92%	[42]
Pt-TiO ₂	UVA-LEDs	Cyclohexane	200 ppm	68%	[43]
Ag-TiO ₂	UV 8W	Toluene	120 ppm	98%	[44]
Cu-TiO ₂	Light bulb 20 W	Benzene	100 ppm	86%	[45]
Mn-TiO ₂	UVC & VUV	Benzene	50 ppm	58%	[46]
Ni-TiO ₂	UVC 10W	p-Xylene	80 ppm	60%	[47]
Fe-TiO ₂	UV light	Toluene	0.1 ppm	63%	[48]
TiO ₂ /MnO ₂	VUV 4 W	Toluene	30 ppm	96%	This work

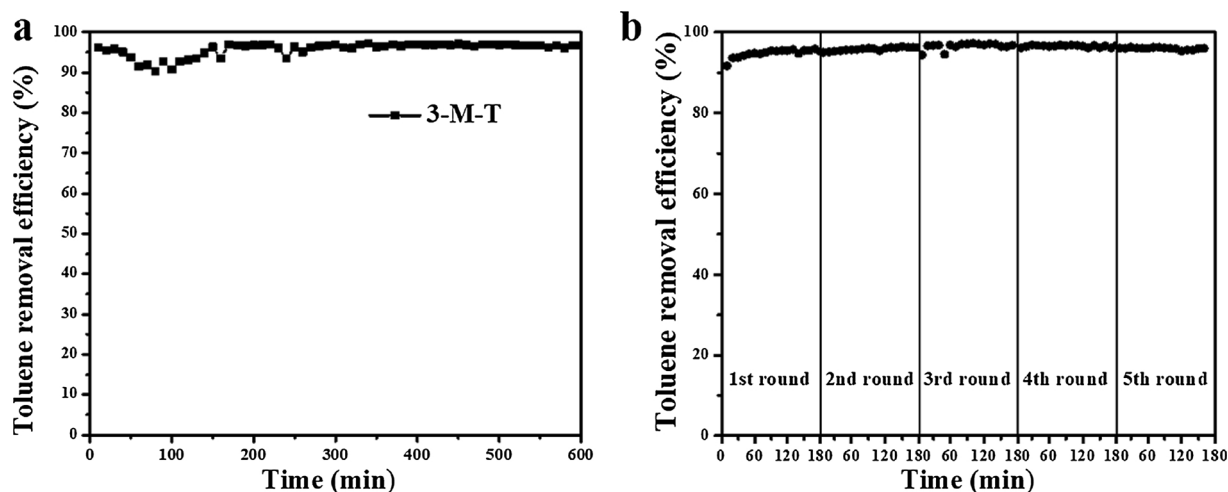


Fig. 10. (a) Stability and (b) regenerated ability of 3-M-T for VUV-PCO degradation of Toluene.

a good electron acceptor and reduced to $\cdot\text{O}_3^-$ radicals, which can prevent the recombination of electron-hole pairs [20,58,59]. Meanwhile, the adsorbed H_2O could be oxidized by the $\cdot\text{O}_3^-$ radicals to form $\cdot\text{OH}$ radicals. Moreover, $\cdot\text{O}_3^-$ radicals can oxidize toluene with a relatively long lifetime according to the literature [60]. On the other hand, the holes left in the valance band (VB) of TiO_2 could oxidize OH^- to form $\cdot\text{OH}$ radicals.

In addition, Zhang et al. indicated that catalytic ozone decomposition is associated with oxygen vacancies (V_O) [61]. First, the oxygen vacancies transfer two electrons to an O atom of an O_3 molecule that bounded on the $\text{TiO}_2/\text{MnO}_2$ heterojunction surface, the O_3 can then be decomposed into an oxygen species (O^-) and a bridging O_2 dimer. The O^- can further react with another O_3 molecule to release an O_2 and a peroxide species (O_2^-). After that, the generated O_2^- reacts with H_2O to generate $\cdot\text{OH}$ radicals for toluene degradation. Finally, the generated O_2^- can be decomposed to release an oxygen molecule, and at the same time, V_O is recovered and can participate in the next cycle to decompose O_3 [62,63]. Therefore, the catalytic decomposition of toluene and O_3 on $\text{TiO}_2/\text{MnO}_2$ NSs involved in the VUV-PCO process occurs via the following two phases:

Phase I:

- $\text{TiO}_2 \xrightarrow{h\nu} e^- + h^+$
- $h^+ + \text{H}_2\text{O} \rightarrow \cdot\text{OH} + \text{H}^+$
- $\text{O}_3 + e^- \rightarrow \cdot\text{O}_3^-$
- $\cdot\text{O}_3^- + \text{H}_2\text{O} \rightarrow \cdot\text{OH} + \text{OH}^- + \text{O}_2$

Phase II:

- $\text{O}_3 + \text{V}_\text{O} \rightarrow \text{O}^- + \text{O}_2$
- $\text{O}_3 + \text{O}^- \rightarrow \text{O}_2^- + \text{O}_2$
- $\text{O}_2^- + \text{H}_2\text{O} \rightarrow 2\cdot\text{OH}$
- $\text{O}_2^- \rightarrow \text{O}_2 + \text{V}_\text{O}$

Overall, under the irradiation of VUV light and taking advantage of the photocatalytic effect of TiO_2 NSs as well as the catalytic ozonation of Mn, the $\text{TiO}_2/\text{MnO}_2$ heterojunction catalyst exhibited multi-functions for efficient toluene degradation and decomposition of all the residual O_3 in the VUV-PCO process.

4. Conclusions

The heterojunction structure of $\text{TiO}_2/\text{MnO}_2$ NSs was developed and tested for toluene degradation in the VUV-PCO process. The toluene removal efficiency can reach 96% and almost all the residual ozone can be removed over the $\text{TiO}_2/\text{MnO}_2$ heterojunction catalyst under the irradiation of VUV light. The high performance of toluene degradation of the catalyst can be attributed to the introduction of MnO_2 to the surface of the TiO_2 NSs, which enhanced the light absorption ability and inhibited the recombination of photo-induced electron-hole pairs, thus, leading to an enhanced photocatalytic activity for toluene degradation. In addition, the catalytic ozonation of Mn can effectively utilize the generated O_3 to improve the efficiency of toluene degradation in the

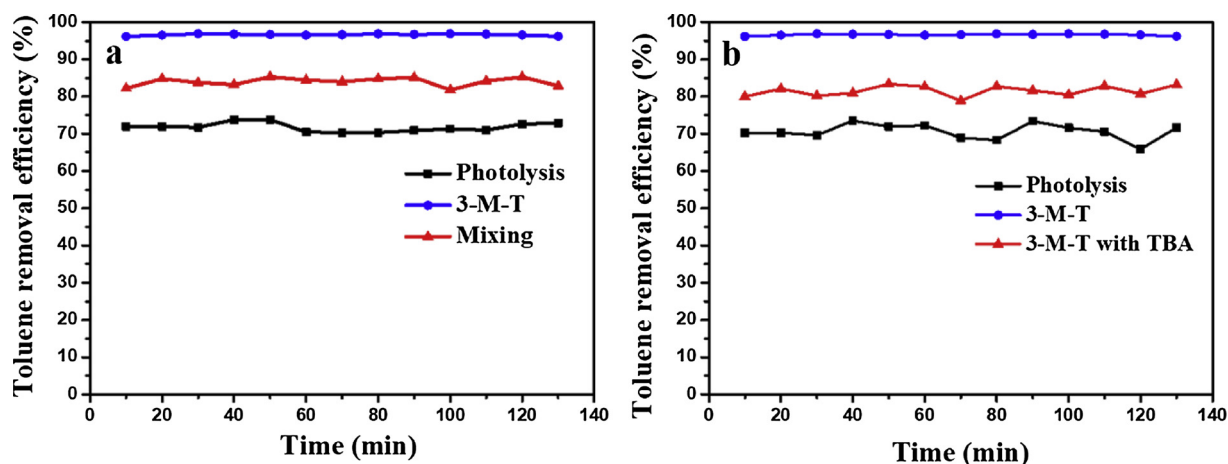


Fig. 11. Toluene degradation performance of the as-prepared samples of (a) TiO_2 NSs mixed with MnO_2 and (b) 3-M-T with TBA.

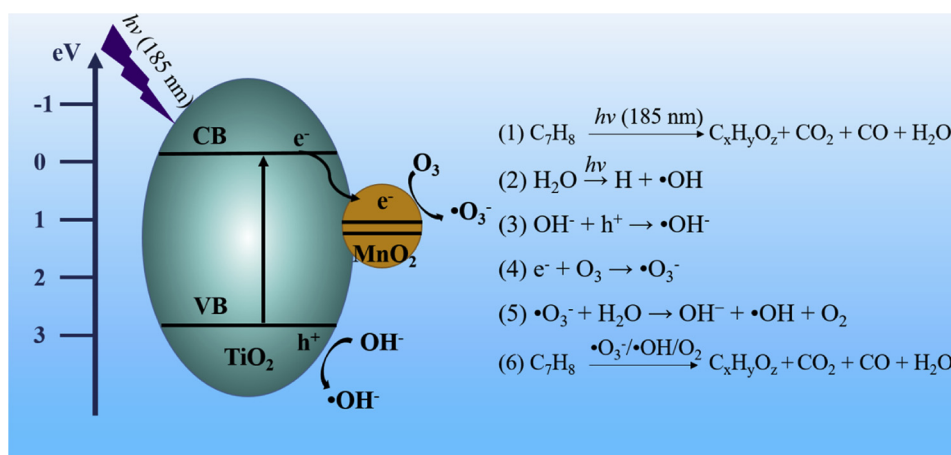


Fig. 12. Schematic diagram of the photocatalytic mechanism of toluene degradation under the VUV-PCO process over $\text{TiO}_2/\text{MnO}_2$ heterojunction catalyst.

VUV-PCO process, as well as the elimination of residual ozone in the system. This work provides a new way of developing photocatalysts with high efficiency for VOCs degradation.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supplementary data

Supplementary material related to this article can be found, in the online version, at doi:<https://doi.org/10.1016/j.apcatb.2019.118034>.

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